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1 **The seascape nursery: a novel spatial approach to identify and**
2 **manage nurseries for coastal marine fauna**

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4 Alternative 1: The seascape nursery: incorporation of dynamic processes to identify and
5 manage nurseries for coastal marine fauna

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7 Alternative 2: Identification and management of nurseries for coastal marine fauna based on a
8 novel seascape approach

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24 Running title: The seascape nursery

25 **Abstract**

26 Coastal marine and estuarine ecosystems are highly productive and serve a nursery function
27 for important fisheries species. They also suffer some of the highest rates of degradation from
28 human impacts of any ecosystems. Identifying and valuing nursery habitats is a critical part of
29 their conservation, but current assessment practices typically take a static approach by
30 considering habitats as individual and homogeneous entities. Here we review current
31 definitions of nursery habitat and propose a novel approach for assigning nursery areas for
32 mobile fauna that incorporates critical ecological habitat linkages. We introduce the term
33 ‘seascape nurseries’ which conceptualizes a nursery as a spatially-explicit seascape consisting
34 of multiple mosaics of habitat patches that are functionally connected. Hotspots of animal
35 abundances/productivity identify the core area of a habitat mosaic, which is spatially
36 constrained by the home ranges of its occupants. Migration pathways connecting such
37 hotspots at larger spatial and temporal scales, through ontogenetic habitat shifts or inshore–
38 offshore migrations, should be identified and incorporated. The proposed approach provides a
39 realistic step forward in the identification and management of critical coastal areas, especially
40 in situations where large habitat units or entire water bodies cannot be protected as a whole
41 due to socio-economic, practical, or other considerations.

42

43 **Keywords** Ecosystem connectivity, juvenile fauna, mangrove, ontogenetic migration, salt
44 marsh, seagrass

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46

47 **Introduction**

48

49 Coastal ecosystems provide a range of valuable ecosystem services, such as fisheries
50 production, protection against coastline erosion, and carbon sequestration (Costanza *et al.*
51 1997). With about 60% of the world's population living within 100 km of the coastline
52 (Vitousek *et al.* 1997) these ecosystems have suffered from rapid degradation (Waycott *et al.*
53 2009). Coastal and estuarine systems are highly productive and important for food security
54 and livelihoods. Where multiple ecosystems are hydrologically and ecologically connected, a
55 key function is the replenishment of offshore populations of commercially and ecologically
56 important species of fish and crustaceans (Beck *et al.* 2001). The nursery function of these
57 systems has received much attention over the last decade but current procedures for
58 identifying and evaluating critical habitats lag our scientific understanding of processes that
59 drive nursery function and productivity. In this perspective we propose a novel approach for
60 delineating nursery areas for mobile fauna, incorporating ecological habitat linkages resulting
61 from animal movements that occur at different spatial and temporal scales.

62 Three lines of research tackle the issue of coastal ecosystem connectivity for marine
63 fauna, but at different conceptual scales. Firstly, the nursery-role hypothesis is mainly focused
64 on identifying the nursery habitats that contribute most to offshore adult populations (Beck *et*
65 *al.* 2001; Nagelkerken 2009). Secondly, ecosystem-connectivity studies have largely
66 attempted to correlate a variety of structural metrics of coastal nursery habitats to catches of
67 offshore fishery stocks (Manson *et al.* 2005). Finally, seascape studies have applied
68 techniques and concepts from landscape ecology to understand what drives the spatial
69 patterning of animal communities in coastal nursery habitats (Sheaves and Johnston 2008;
70 Boström *et al.* 2011). While each of these research directions has received increasing attention
71 in the last decade or two, lack of integration between them has led to gaps in the development
72 of appropriate conservation and management strategies.

73 The nursery-role and ecosystem-connectivity approaches typically consider critical
74 habitats as individual, homogeneous entities. This potentially forces managers faced with
75 conflicting objectives for conservation and alternative uses to evaluate and then trade off
76 entire habitats against one another when determining priorities (Weinstein 2008). Moreover,
77 protected areas with fixed boundaries are ineffective in protecting moving or transient species
78 (Rayfield *et al.* 2008). The seascape-ecology approach points to a different solution, based on
79 mosaics of habitat patches at smaller spatial scales (Simenstad *et al.* 2000). The spatial
80 characteristics of habitat patches play an important role in structuring associated animal
81 communities, but typically are not considered in assessments of nursery value, leaving a
82 critical knowledge and conservation gap (Beck *et al.* 2001; Adams *et al.* 2006; Boström *et al.*
83 2011).

84 Previous attempts to define marine nurseries have provided an important, but relatively
85 basic, framework for the identification of nursery habitats. These approaches are static in that
86 they do not indicate how to specifically incorporate dynamic processes, such as ontogenetic
87 habitat shifts, animal movement, and spatially-explicit usage of habitat patches and corridors
88 within seascapes. This static, single-habitat approach potentially leads to incomplete or
89 incorrect identification of critical habitats. The aim of this paper is to take a more holistic
90 approach in identifying nurseries. We view a nursery as a spatially-explicit seascape unit
91 (rather than a habitat unit) consisting of functionally-connected mosaics of habitats
92 incorporating ecological processes driven by animal behaviour, and define this as the
93 'seascape nursery'.

94

95

96 **Review of nursery function definitions**

97

98 Early designations of nursery habitats simply referred to habitats with high densities of
99 juvenile animals. Beck *et al.* (2001) greatly improved the definition by arguing that a nursery
100 is a habitat contributing a higher than average biomass of juveniles per unit area to the adult
101 population than other habitats, resulting from higher densities, higher growth, lower mortality
102 and/or greater movement. However, this approach under-appreciates juvenile habitats that
103 have a large surface area but low density of organisms, even though their overall contribution
104 to the adult population might be larger. Therefore, Dahlgren *et al.* (2006) suggested that
105 identification of nurseries should be based on their total contribution to the adult population.
106 This was criticized as an approach that failed to consider the importance of dynamic processes
107 that underpin nursery function (Sheaves *et al.* 2006), but no specific solutions were offered
108 (Layman *et al.* 2006). While some studies (e.g. Beck *et al.* 2001; Adams *et al.* 2006) have
109 covered important factors that regulate nursery value, no significant steps towards a more
110 comprehensive and realistic method for the identification of nurseries have occurred. Clearly,
111 managing a nursery habitat as a whole unit will not be effective without considering the
112 sequence of habitats that are used throughout ontogeny, while other aspects of nursery
113 habitats (e.g. movement corridors, density hot-spots) should be considered to conserve the
114 most productive and important habitat patches within nursery habitats. Some of these aspects
115 have been briefly mentioned in previous studies (Beck *et al.* 2001; Adams *et al.* 2006), but a
116 framework of how to address these issues is still lacking. In the present study we propose a
117 potential framework to enhance identification and conservation of nurseries.

118 We concur with the current view that the value of nurseries (as defined by Beck *et al.*
119 2001) relates to their ultimate contribution to the support of populations. However, we move
120 beyond the approaches that identify nurseries as static habitat units, and provide a perspective
121 on how advances in seascape ecology can enhance designation and valuation of nursery
122 habitats for animals that use inshore habitats before migrating offshore (“ontogenetic

123 shifters”; Adams *et al.* 2006). Like previous efforts, our goal is to improve the management
124 and conservation of critical nursery habitats. Here we build on those efforts to gain an
125 improved measure for nursery habitat designation that captures critical processes and habitat
126 linkages that underpin nursery function and might otherwise be missed by earlier approaches.

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129 **Early-juvenile population bottlenecks: identifying critical settlement habitats**

130

131 Searching for preferred habitat while in the water column is risky and therefore settlement-
132 stage larvae often occupy the first-encountered suitable habitat when entering estuaries or
133 lagoons from the open ocean (Grol *et al.* 2011), with subsequent shifts to other habitats in a
134 step-wise pattern (Cocheret de la Morinière *et al.* 2002). Less structurally complex habitats
135 such as sand patches, macroalgal clumps or dead coral rubble may function as important
136 settlement habitats (Dahlgren and Eggleston 2000), but are often disregarded in their value for
137 settling larvae. The identity of transient settlement habitats is unknown for many species, they
138 may be occupied only briefly, yet they may well form population bottlenecks for early post-
139 settlement stages (Fodrie *et al.* 2009). They are easily missed because of the small sizes at
140 which juveniles occupy these transient habitats and because of the relatively short duration of
141 occupancy. However, many species settle from the plankton during specific seasons of the
142 year, and field surveys should be performed during these seasons to identify important
143 settlement areas. We specifically recommend that these often-missed first-stage habitats be
144 considered in the seascape nursery concept.

145

146

147 **Habitat connectivity: predictable diel, tidal, and ontogenetic habitat shifts**

148

149 Few species are confined to a single nursery habitat (Nagelkerken 2007). Seascape studies
150 have shown that many animals utilize a mosaic of habitats on a daily basis (Boström *et al.*
151 2011). Mobile animals connect adjoining habitats through tidal, shelter-seeking, or foraging
152 movements (Hammerschlag *et al.* 2010; Igulu *et al.* 2013; Olds *et al.* 2013; Baker *et al.* in
153 press). These migrations are highly predictable in timing and routes followed (Krumme
154 2009), to such extent that some predators in nursery areas have adapted their behaviour to
155 coincide with these migrations (Helfman 1986). Animals pass through non-nursery habitats
156 on a regular basis while moving between patches of core habitat in search of food or shelter
157 (Hitt *et al.* 2011). These movements usually occur within a specified home range around the
158 core area of their shelter sites (Farmer and Ault 2011), which are often located near to
159 structurally-complex habitats (Verweij and Nagelkerken 2007). Species often show homing
160 behaviour to such shelter sites, which may persist over periods of weeks to months (Helfman
161 *et al.* 1982). On longer time-scales, many species show ontogenetic shifts among habitats
162 because of changing resource needs (e.g. food, shelter) as well as altered predation risk during
163 different life stages (Dahlgren and Eggleston 2000; Kimirei *et al.* 2013b). Due to strong
164 connectivity among habitat patches, assigning single nursery habitats disregards the role that
165 earlier life-stage habitats or adjoining (feeding/shelter) habitats play in the population
166 dynamics and ultimate stock replenishment of nursery species.

167

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169 **The seascape mosaic: hotspots of animal abundances and productivity**

170

171 Spatially-explicit use of patches within nursery habitats typically has not been quantified in
172 relation to nursery function. In contrast, landscape-focused studies have demonstrated

173 consistent and predictable animal density or productivity ‘hotspots’ in relation to spatial
174 position within the seascape, for example based on: 1. distance to estuary mouth (Bell *et al.*
175 1988), 2. distance to feeding areas (Pittman *et al.* 2007), 3. proximity to high-volume tidal
176 channels that supply larvae (Ford *et al.* 2010), 4. density of creek edges within marshes
177 (Kneib 2003), 5. presence and type of adjacent habitats (Nagelkerken *et al.* 2001), or 6.
178 specific salinity regimes representative of transitional areas between rivers and estuaries
179 (Wasserman and Strydom 2011). Furthermore, habitat transition areas are specific zones
180 within coastal seascapes that often have greater densities of organisms than areas further from
181 edges (Dorenbosch *et al.* 2005). In many cases the Beck *et al.* (2001) and Dahlgren *et al.*
182 (2006) approaches may well identify the broad nursery habitat(s) used by a population, but
183 miss critical mosaics of habitat patches in the seascape that underpin nursery function
184 (Sheaves 2009).

185

186

187 **Ecosystem corridors: highways connecting nurseries to adult populations**

188

189 The last stage of nursery habitat occupancy during which organisms undertake their final
190 migration to deeper or offshore waters to join the adult population is poorly known
191 (Gillanders *et al.* 2003), but telemetry studies suggest that it can occur over short periods
192 ranging from a few hours to days (Luo *et al.* 2009). Specific routes within estuaries or lagoons
193 may act as preferred corridors that lower predation risk, span the shortest distance to reach
194 deeper water, or facilitate tidally-enhanced movements due to specific local hydrology
195 (Zollner and Lima 1999). Some studies have indicated the importance of continuous habitat
196 edges (Hitt *et al.* 2011) or unvegetated strips within continuous seagrass beds as corridors
197 (Boström *et al.* 2006), but extensive open shallow areas normally act as barriers for

198 movement (Turgeon *et al.* 2010). In intertidal areas with extensive sand or mud flats, animals
199 will often be funnelled to subtidal habitats through narrow tidal channels. From there on, fish
200 move to offshore waters by navigating through corridors such as deep channels, through
201 narrow bay mouths, or through open spaces among sand banks, islets and other types of
202 natural barriers situated at the ocean side of river deltas, estuaries and lagoons (e.g. Verweij *et*
203 *al.* 2007; Luo *et al.* 2009). Incorporation of migration corridors and their temporal usage
204 patterns is a critical consideration for the seascape nursery concept.

205

206

207 **The seascape nursery: combining nursery-function and seascape-ecology concepts**

208

209 Existing approaches to nursery habitat evaluation tend to give more weight to final juvenile
210 stages prior to emigration to offshore adult stocks. Linkages among habitats that affect the
211 critical growth and survival of earlier stages therefore tend to be underplayed. We suggest that
212 the seascape nursery approach incorporates more fully those earlier stages. The importance of
213 our approach is demonstrated in the following example for fishes with a complex life cycle.
214 Consider a micro-tidal seascape (Fig. 1a) where fish settle largely in first-encountered, non-
215 core habitats like coral rubble areas along edges of tidal channels or at bay mouths,
216 subsequently progress to seagrass beds, then switch to mangroves, and finally occupy hard-
217 bottom patch reefs or rocky areas, before moving to offshore reefs (example from
218 Nagelkerken *et al.* 2000 and Grol *et al.* 2011). In this example, individuals are also found in
219 other habitats, but those described above are where highest fish aggregation or production
220 occurs. During seagrass and rubble occupancy small juveniles feed and shelter in the same
221 habitat to reduce predation, but at larger sizes they use mangroves or patch reefs for shelter
222 and show a diel or tidal migration to nearby seagrass beds to feed (Verweij *et al.* 2006).

223 During these movements, they need to move from one feeding patch to another and pass
224 through secondary habitats, such as algal beds and sand patches, which do not play an
225 important role for feeding or as shelter but are part of their home range (see concentric circles
226 in Fig. 1).

227 In the above example, the extensive seagrass beds provide the largest overall contribution
228 to the adult populations (e.g. Verweij *et al.* 2008) and would be identified as the main nursery
229 habitats using the Dahlgren *et al.* (2006) approach. In contrast, expressed as a contribution per
230 unit area the importance of seagrass beds with large surface area would typically be lower
231 compared to other habitats with smaller surface areas where crowding of animals occurs, like
232 mangrove stands and coral patches. Based on the Beck *et al.* (2001) approach such habitats
233 that contribute most per unit area could be designated as nursery habitats even though their
234 overall production might not be large. This could in practice lead to a debate about whether
235 mangroves versus seagrass beds should be managed, what proportion of their total surface
236 area should be conserved, and which areas within the estuary or lagoon should be managed,
237 especially in cases of high-usage or exploitation by multiple stakeholders. The seascape
238 nursery would provide a more realistic approach to this problem by revealing that (Fig. 1): 1.
239 transient settlement areas should be conserved, because without these there is no recruitment
240 to 'nursery' habitats, 2. within the seascape there are principle areas (habitat mosaics),
241 constrained by animal home ranges, that attract higher densities of mobile organisms and
242 which are more productive than other areas, providing a management tool to prioritize areas
243 of conservation, 3. successive essential life-stage habitats should be conserved as impacts on
244 one habitat affect productivity in habitats occupied during later life-stages, 4. without
245 conserving migration routes that connect different animal hotspots during ontogeny or that
246 facilitate movement from nurseries to offshore populations, nurseries could experience a

247 switch from acting as sources to becoming juvenile sinks. A similar example from a meso-
248 tidal salt marsh system is provided in Figure 1B.

249 Not all species show a complex life cycle such as described above. Nevertheless, it is a
250 common observation for a multitude of species that tidal channels are favoured for movement
251 through shallow areas, that animal abundances are highly correlated with spatial position
252 within coastal habitats (e.g. driven by salinity or turbidity gradients), and that animals
253 regularly perform diel or tidal movements (Whaley *et al.* 2007; Krumme 2009; Turgeon *et al.*
254 2010). So even for species with a relatively simple life cycle, in terms of habitat use, previous
255 approaches fail to incorporate several important dynamic processes other than ontogenetic
256 habitat shifts.

257

258

259 **Practical steps to seascape nursery analysis**

260

261 While there is no single best approach to identify the precise mosaic of habitats most essential
262 during the juvenile stages of animals in coastal marine environments, it is crucial to recognize
263 the importance of a mosaic of contributing habitats and their linkages. Here we outline the
264 practical steps that can help improve on earlier approaches for identification and evaluation of
265 nursery habitat and ultimately lead to more successful protection and management of nursery
266 function. The order and relative importance of these steps will vary depending on specific
267 situations.

268 Step 1: Following Beck *et al.* (2001), identify the relative contribution to adult
269 populations of all juvenile habitats at whatever scale they can be identified, e.g. using
270 approaches such as otolith microchemistry (Gillanders and Kingsford 1996; Verweij *et al.*
271 2008). This will typically be at a coarser scale than relevant to management objectives (e.g.

272 whole estuary or whole habitat unit) and fail to identify linkages across the seascape. We
273 therefore recommend subsequent work to identify the smaller-scale patches within each
274 broad-scale nursery that contribute most to the overall population replenishment by that
275 nursery habitat. This will likely, but not necessarily overlap with density hotspots of juvenile
276 animals during their inactive as well as active period (e.g. Ford *et al.* 2010), which can be
277 identified through field surveys. Identification of specific patches that contribute most to the
278 overall production of a nursery habitat is more challenging, but techniques such as stable
279 isotope analysis of muscle tissue, internal and external artificial tags, or genetic and chemical
280 markers can provide the necessary finer-scale information (Gillanders 2009; Kimirei *et al.*
281 2013a), as well as provide an answer to how this contribution may vary over time (see e.g.
282 Kraus and Secor 2005).

283 Step 2: Known (from the literature) or field-acquired (through tagging studies) home
284 range sizes may then be projected onto the identified highest-productivity density-hotspots to
285 establish the effective area that is used as a juvenile habitat (the habitat mosaic). The home
286 range includes the seascape that is most used on a daily basis for activities such as sheltering
287 and foraging. Home range sizes around hotspots of animal abundances could be considered at
288 decreasing levels of importance (see Fig. 1). Using radii of these dimensions should prove to
289 be a more effective way to manage nursery mosaics than a static approach of single complete
290 habitats because it uses broader information on critical habitat use. While tagging juvenile
291 animals is difficult and movement ranges can differ considerably among species and within
292 habitats, home range size is often a function of body size (Kramer and Chapman 1999) and
293 juveniles of most demersal species show high site fidelity and restrict their movements to
294 distances of no more than a few 100s m from their preferred shelter sites (Tupper 2007;
295 Nagelkerken *et al.* 2008). Home ranges are larger in cases where animals occupy macrotidal

296 habitats, but also in this case fidelity has been shown to high-tide and low-tide habitat
297 components (Dorenbosch *et al.* 2004; Hering *et al.* 2010).

298 Step 3: Patterns of ontogenetic habitat shifts should be identified for animals that occupy
299 the above high-productivity hotspots, so that other habitat patches that are previously or
300 subsequently occupied are included in the designation of effective nursery mosaic (Fig. 1).
301 This is based on the principle that patches that contribute most to adult populations can only
302 sustain this productivity as a result of habitat linkages through ontogeny. Approaches such as
303 following the progression of cohorts (abundances and sizes of organisms) in multiple juvenile
304 habitats can identify which habitats are most likely to play a key role in provisioning recruits
305 to next life-stage habitats (e.g. Fodrie *et al.* 2009). A critical consideration in this is to identify
306 primary settlement areas where early life stages occur, typically at sizes at which they have
307 not been included in field surveys.

308 Step 4: Primary migration routes should be identified (e.g. using telemetry or
309 conventional tagging) that connect animal production hotspots across different spatio-
310 temporal scales. This includes corridors that facilitate animal movement from one habitat
311 mosaic to another through ontogeny, as well as from the seascape nursery to offshore waters
312 (Fig. 1). Migration highways are likely to overlap among species based on the same
313 advantages that they provide for a suite of species, like structure-rich corridors that facilitate
314 movement under lowered predation risk (Gilliam and Fraser 2001). In deep-water estuaries
315 and lagoons such migration corridors might be less evident or relevant than in shallower
316 ecosystems dominated by extensive mud or sand-flats. However, due to the geomorphology
317 of many inshore water bodies around the world, animals still need to pass through bay
318 mouths, openings between barrier islets, or through deeper tidal channels to reach offshore
319 waters. As such, these areas should be given high conservation importance as they maintain
320 connectivity among inshore and offshore ecosystems.

321 We have attempted to present an improved framework to identify nurseries for
322 management purposes that we believe will provide an acceptable level of accuracy for a wide
323 range of species in a variety of coastal marine ecosystems. Our approach does not provide a
324 single, best solution for multi-species management, as different groups of species may occupy
325 different combinations of habitats or different areas of estuaries and lagoons. As is the case
326 for previous approaches of nursery identification, trade-offs need to be made in terms of
327 which species and which areas receive most consideration in terms of conservation or
328 management. While for some systems with few, highly abundant fishery species and just one
329 or two habitat types, a coarse approach such as that of Dahlgren *et al.* (2006) and Beck *et al.*
330 (2001) may provide a reasonable amount of information for management purposes, there are
331 many other systems and a multitude of (commercial and keystone) species where such an
332 approach is likely to fail. The seascape nursery approach adds more realism to the
333 identification of core juvenile areas within these systems by incorporating spatio-temporal
334 drivers of animal habitat use. The intention is to achieve a practical advance for the
335 conservation and management of inshore coastal areas that are highly productive for coastal
336 fisheries but also prone to high levels of competing demands and degradation through human
337 activities. We also recommend consideration of more challenging, dynamic management
338 approaches such as mobile protected areas that follow movements of key species across their
339 landscape (Bull *et al.* 2013).

340

341

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345

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347 **References**

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532 **Figure legend**

533

534 **Figure 1** (a) Example as described in the text of a seascape nursery located in a clear-water,
535 micro-tidal lagoon supporting a variety of habitat types; the seascape nursery consists of
536 several habitat mosaics connected through diel and ontogenetic movements. (b) Example of a
537 seascape nursery for penaeid shrimps in a turbid, meso-tidal salt marsh estuary. This specific
538 case study refers to coastal salt marsh ecosystems of the northern Gulf of Mexico which are
539 considered critical in the support of highly productive shrimp fisheries (Turner 1977). Adults
540 spawn offshore and post-larvae recruit to shallow habitats in the marsh complex of coastal
541 bays and estuaries where conditions are favourable (salinity, temperature, food availability)
542 (Rozas and Minello 2011). There is a staged ontogenetic progression of juveniles from the
543 marsh complex to open bays, and subsequent migration to join adult stocks offshore (Lindner
544 and Cook 1970). Although represented as circles for consistency of presentation, a narrow
545 strip at the vegetation–open water interface represents a density hotspot for juvenile shrimp
546 within the marsh complex (Minello *et al.* 2008). Image credits: Kate Moore, Jane Thomas,
547 Tracey Saxby and Diana Kleine (IAN Image Library – ian.umces.edu/imagelibrary) and Nina
548 McLean (James Cook University).

Figure 1

